

1 **Heavy metal and trace element concentrations in the blood of scalloped**
2 **hammerhead sharks (*Sphyrna lewini*) from La Paz Bay, México**

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41 **Abstract:**

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3 43 Sharks are particularly susceptible to bioaccumulation due to their life history characteristics
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5 44 and trophic position within marine ecosystems. Despite this, studies of bioaccumulation cover
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7 45 only a small proportion of extant species. In this study we report concentrations of trace
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9 46 elements and heavy metals in blood samples of *S. lewini* for the first time. We report high
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11 47 concentrations of several trace elements and heavy metals, with concentrations of some
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13 48 elements exceeding the limit determined safe for human consumption. High elemental
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15 49 concentrations may reflect biochemical differences between blood plasma and other tissues;
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17 50 however, they may also be symptomatic of high levels of exposure triggered by anthropogenic
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19 51 activities. We also provide evidence of elemental accumulation through ontogeny, the nature
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21 52 of which differs from that previously reported. Ultimately, this baseline study increases our
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23 53 understanding of interspecific and intraspecific variation in bioaccumulation and
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25 54 ecotoxicology in elasmobranchs which may prove important in ensuring adequate
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27 55 management.
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39 58 **Key words:** Elasmobranchii, bioaccumulation, turnover rate, blood plasma, conservation
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67 **Introduction:**

68 Pollution and bioaccumulation are two of the most significant anthropogenic factors affecting
69 marine communities (Todd et al., 2010; Ley-Quiñónez et al., 2013; Boldrocchi et al., 2020).
70 Whilst lethal effects have been reported in a number of taxa (Martin and Holdich, 1986) sub-
71 lethal bioaccumulation of both elemental and organic contaminants is arguably an issue of even
72 greater importance, given both the potential for trophic magnification (Gelsleichter et al., 2006;
73 Gelsleichter & Walker, 2010; Storelli & Marcotrigiano, 2001) and difficulties in determining
74 the true extent of bioaccumulation within any given species (Burkhard et al., 2012; Caurant et
75 al., 1999; Rodríguez-Romero et al., 2021). Chronic, or intermittent exposure to these
76 contaminants can significantly affect physiology, cellular biology, and behavior of living
77 organisms (Gelsleichter et al., 2006; Jepson et al., 2005; Thophon et al., 2004; Ylitalo et al.,
78 2005;), with these effects varying between and within species (Boldrocchi et al., 2019).
79 Crucially, these effects can have pervasive and unpredictable long-term consequences for
80 ecological communities as a result of interactions between taxa and the biogeochemical
81 components of their environment (Moiseenko, 2017). Moreover, the chemical behavior of trace
82 elements is likely to change in the face of anthropogenic climate change (Alves et al., 2023;
83 Cabral et al., 2019; Cao et al., 2015; Eagles-Smith et al., 2014; Rodríguez-Romero et al., 2021;)
84 and thus there is no guarantee that previous investigations of trace element bioaccumulation
85 are representative of contemporaneous trends, or those that may be observed in the future. It
86 should also be noted that marine bioaccumulation represents a potential health hazard for
87 human populations (Akhbarizadeh et al., 2018; Cara et al., 2022), particularly in coastal
88 communities relying on artisanal fisheries for a substantial proportion of their diet (Cartamil et
89 al., 2011; Galvan-Magaña et al., 2013; Storelli et al., 2003; Vázquez-Hurtado et al., 2010;). For
90 these reasons, establishing the extent to which bioaccumulation is affecting marine organisms
91 (at all trophic positions and in all ecosystems) is of critical importance.

92 Elasmobranchs (sharks, rays, and skates) are particularly susceptible to
93 bioaccumulation as a result of their *k-selected* life-history parameters (Castro, 1993; Mull et
94 al., 2012). Many elasmobranch taxa also occupy higher trophic levels (Maz-Courrau et al.,
95 2011; Wosnick et al., 2021), a factor which has been shown to influence the bioaccumulation
96 of many marine pollutants (Gelsleichter and Walker, 2010; Maz-Courrau et al., 2011; Pancaldi
97 et al., 2019; Storelli et al., 1998, 2003). Bioaccumulation, its potential biological consequences,
98 and putative adaptations to counteract them have been studied in several elasmobranch taxa
99 (Escobar-Sánchez et al., 2010; Pancaldi et al., 2021; Wosnick et al., 2021), however such
100 studies only account for a relatively small proportion of total elasmobranch diversity (Lee et
101 al., 2015; Mull et al., 2012). Assessing the nature and potential consequences of trace element
102 bioaccumulation in elasmobranch taxa at the population level is important given their intrinsic
103 vulnerability to anthropogenic stressors such as overfishing (Dulvy et al., 2017; Walker, 1998),
104 and the potential ecosystem-wide consequences of predator declines (Palkovacs et al., 2011;
105 Parsons, 1992; Polis et al., 2000).

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107 The Scalloped Hammerhead shark (*Sphyrna lewini*) is a large-bodied, migratory
108 carcharhiniform shark, distributed worldwide in tropical and temperate oceans, where they
109 inhabit both coastal and pelagic zones at depths of up to 275 meters (Compagno, 2011; Gulak
110 et al., 2015; Moore & Gates, 2015). *S. lewini* is known to be in decline globally (Hayes et al.,
111 2009), and has been listed as critically endangered by the International Union for the
112 Conservation of Nature (IUCN) since 2018, predominantly due to overexploitation in longline
113 fisheries (Aldana-Moreno et al., 2020) and low rebound potential (Pancoureau et al., 2018,
114 2021). Heavy metal and trace element concentrations have been studied previously in the liver
115 and muscle tissues of *S. lewini*, however these tissues are thought to have a slow turnover rate
116 relative to blood plasma, such that heavy metal and trace element concentrations do not

117 necessarily reflect the current biogeochemical environment at the location in question (Kim et
118 al., 2012; Whitehead et al., 2020). Such studies generally report heavy metal and trace element
119 concentrations lower than those which would present risk to human consumers (Bergés-
120 Tiznado et al., 2015, 2021; Ruelas-Inzunza et al., 2020). Some biological consequences marine
121 pollution for *S. lewini* have also been considered (Boswell, 2015), although the extent to which
122 these results are indicative of *S. lewini* populations globally remains poorly constrained. In this
123 study we report concentrations of trace elements and heavy metals in blood samples of *S. lewini*
124 for the first time. We consider the concentration of various trace elements and heavy metals in
125 isolation (including implications for human health and ecology), as well as potential
126 relationships between elemental concentrations and key life history parameters (size and sex).
127 Finally, we consider the extent to which *S. lewini* may have evolved detoxification mechanisms
128 to avoid deleterious consequences of bioaccumulation. This study will increase our
129 understanding of interspecific and intraspecific variation in bioaccumulation and
130 ecotoxicology in elasmobranchs. We consider potential ecological and biological
131 consequences of marine pollution in this system and suggest reasonable directions for future
132 work and conservation action.

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134 **MATERIALS AND METHODS**

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136 **Ethics statement:**

137 Data collection and analysis procedures in this study complied with national animal welfare
138 laws; guidelines and policies and was authorised by Mexican wildlife authorities under the
139 permit PPF/DGOPA-024/20 provided by the Comisión Nacional de Acuacultura y Pesca
140 (CONAPESCA). Participants of this study neither promoted nor encouraged the harvesting of
141 sharks and all samples were collected with the consent of the artisanal fishing communities.

142 **Sample collection:**

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2 143 Blood samples were obtained from a total of 126 *S. lewini* individuals between December 1st
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5 144 2020 and February 8th 2021 at the El Saladito fish camp on the west coast of La Paz Bay, Baja
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7 145 California Sur, Mexico (Figure 1). The highest number of individuals were registered in the
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10 146 month of December (n=103) accounting for a total of 81.7% of the sampled individuals,
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12 147 followed by February (n=13) with 10.3% and January (n=10) with 7.9%. Approximately 5 ml
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14 148 of whole blood was collected directly from the heart of each shark immediately after landing
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17 149 and stored in sterilized tubes at -40°C for further analyses. In addition, the following biological
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19 150 information was collected from each individual: total body length, measured from the tip of
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22 151 the snout to the apex of the upper lobe of the caudal fin (TL), and sex, determined by the
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24 152 presence or absence of the male intromittent organs (claspers). All sharks examined within this
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27 153 study were juvenile, as they were smaller than the estimated size at maturity for the species
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29 154 (Compagno, 2005). Size distribution ranged from 81 cm to a maximum size of 166 cm in TL,
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32 155 with a mean size of 108.9 cm ± 17.3 cm. Two size classes were defined based on TL with
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34 156 sharks < 100 cm considered ‘immature’, whereas individuals with TL ranging between 100
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36 157 and 180 cm were considered ‘subadult’.

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41 159 **Sample analysis:**

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44 160 To remove excess moisture, blood samples were lyophilized using a LABCONCO freeze dry
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46 161 system at a constant temperature of -50°C for a total of 48 hrs at the Centro Interdisciplinario
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49 162 de Ciencias Marinas (CICIMAR) in La Paz, Mexico. Upon completion of the drying process,
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51 163 all blood samples were ground into a homogeneous powder using agate mortar and pestle,
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54 164 following the approach of Escobar-Sanchez et al. (2010). Subsequent analyses were performed
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56 165 in the university of Milano-Bicocca, Milan, Italy. Samples were hot concentrated by acid
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59 166 digestion in closed containers in specifically designed mineralizer Anton Paar Multiwave 5000

167 microwave ovens. Approximately 0.3 g of each sample was then digested in Teflon vials with
168 5 mL of Nitric Acid - HNO₃ (65%) and 1 mL of Hydrogen Peroxide - H₂O₂ (40%). Following
169 acid digestion, each sample was transferred to a 50 mL plastic tube and brought to 20 mL by
170 adding 14 mL of Milli Q water, in order to lower the acid concentration and avoid compromise
171 of the instrumental reading. Trace elements analyses of Chromium (Cr), Cadmium (Cd),
172 Copper (Cu), Iron (Fe), Nickel (Ni), Selenium (Se), and Zinc (Zn) were performed using an
173 Inductively Coupled Plasma Optical Emission Spectrometry (ICP-OES Optima 7000 DV
174 PerkinElmer) with the Software control WinLab32, following the protocol of Pancaldi et al.
175 (2019). Lead (Pb) and Arsenic (As) were analyzed with the Atomic Absorption
176 Spectrophotometry with Graphite Furnace (GFAAS) PerkinElmer Analyst 600, while Mercury
177 (Hg) was analyzed using AMA254 (Advanced Mercury Analyzer), using approximately 50 mg
178 of each sample inside the furnace. The concentration of the analyte in the sample solution is
179 expressed in ppb (µg/L) using the formula

$$C = (A - b) * V * (d/m)$$

where:

C is the concentration of the analyte in the sample, expressed in µg/L (ppb); A is the analyte
concentration in the sample solution expressed in µg/L (ppb) obtained from the calibration
curve by instrumental reading; b is the reference blank samples for control; V is the total
extracting volume (20 mL); d is the dilution factor; and m the mass of the sample. All
concentrations obtained from the analyses were expressed as µg/g (= ppm; = mg/kg) unless
stated otherwise.

192 **Statistical analyses:**

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2 193 Raw data obtained from the analyses were visualized cleaned before all analyses. Results have
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5 194 been reported as mean \pm standard deviation, unless stated otherwise. Elemental concentrations
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7 195 were compared to those permitted for human consumption by the Food and Drug
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10 196 Administration and the World Health Organization (FAO Legal Notice No 66/2003). The
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12 197 assumption of normality was violated when performing Kolmogorov–Smirnov tests upon the
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14 198 resulting dataset; therefore, nonparametric statistical tests (Mann Whitney and Spearman’s rho)
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17 199 were used to compare the heavy metal and trace elements concentrations between size classes
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19 200 and sexes. The ratio of Selenium (Se) to Mercury (Hg) was evaluated to investigate potential
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22 201 detoxification mechanisms operating in *S. lewini* (Ralston *et al.*, 2007; Ralston & Raymond,
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24 202 2010). All statistical analyses were performed using IBM SPSS 28 Software (IBM SPSS 28,
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27 203 New York, US). Whilst trace element concentrations were established for all 126 individuals,
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29 204 heavy metal concentrations were only investigated in 55 individuals due to an insufficient
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31 205 volume of blood being collected from sampled sharks.
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36 207 **RESULTS**

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41 209 **Elemental concentrations:**

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44 210 Besides Iron (Fe), Arsenic (As) was clearly the element with the highest concentration (16.82
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46 211 ug/g), followed by Nickel (Ni) and Zinc (Zn) with concentration of 5.7 and 4.63 ug/g
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49 212 respectively (Figure 2). Iron (Fe) was expected to be the element present in greatest abundance
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51 213 due to its fundamental biological role in blood. Lead (Pb) was excluded from statistical
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54 214 analyses as it was present in concentrations below the detection limit of the Atomic Absorption
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56 215 Spectrophotometer and Graphite Furnace (GFAAS; dl = 0.05 μ g/L).
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217 Cadmium (Cd) concentration was higher than the edible limit set by the FAO in 40 of the 55
218 individuals tested, while Selenium (Se) in 18 of the 55 individuals tested (Table 1). Moreover,
219 a single individual was found with Mercury (Hg) concentration over the edible limit set by the
220 FAO (Table 1). No other elements were found to occur with concentrations above those
221 considered to be edible by the FAO.

222 223 **Elemental concentrations and size/sex:**

224 Analysis of mean elemental concentrations utilizing both Mann-Whitney and Spearman's Rho
225 tests revealed a statistically significant difference between immature and subadult individuals
226 for the elements Iron (Fe; Mann-Whitney test $p = 0.017$; Spearman's rho Test, Fe: $\rho = 0.330$,
227 $p = 0.016$; Figure 3a) and Chromium (Cr; Mann-Whitney test, $p = 0.016$; Spearman's rho Test,
228 Cr: $\rho = 0.327$, $p = 0.015$; Figure 3a, Table 2). Whilst the concentration of all elements other
229 than Nickel (Ni) appeared to be higher in premature than immature individuals, no additional
230 significant differences between life stages were found ($p > 0.05$). The Spearman's Rho test was
231 also used to test for a relationship between elemental concentration and TL. Significant positive
232 correlations with TL were found for Zinc (Zn) ($p = 0.012$), Iron (Fe) ($p = 0.013$), Chromium (Cr)
233 ($p < 0.001$) and Copper (Cu) ($p = 0.02$), whereas no significant relationship was recovered
234 between TL and the concentration of any other element (Table 3). Whilst the concentrations of
235 all elements appeared to be higher in males than females, neither Mann-Whitney nor
236 Spearman's rho tests recovered evidence of significant relationships between sex and elemental
237 concentration ($p > 0.05$, Figure.3b).

238 239 **Se: Hg ratio:**

240 The molar ratio between Se and Hg could be calculated for only 34 individuals as in the
241 remaining individuals selenium (Se) concentration was below the detection limit of the

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242 apparatus used. Se:Hg ratio was found to be greater than 1 in all but one individual, as the
243 concentration of selenium (Se) exceeded that of mercury (Hg). The lowest ratio recorded was
244 0.55:1, whereas the greatest was 91.98:1 (Figure 4). No relationship between Se:Hg ratio and
245 TL was found (Mann-Whitney test, $p > 0.05$).

247 **DISCUSSION**

249 **Elemental concentrations in *S. lewini*: implications for ecology and human health**

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251 This study is the first assessment of heavy metals and trace elements concentrations from *S.*
252 *lewini* blood plasma. Iron (Fe) was found to be the trace element in greatest abundance (Table
253 1), with a mean concentration comparable to that found in other elasmobranch taxa (Boldrocchi
254 et al., 2019; Merly et al., 2019). This result was entirely expected given the vital role of Fe in
255 blood plasma as a component of hemoglobin (Wells, 1999).

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257 Besides iron, arsenic (As) was the element found in greatest abundance (Table 1), mirroring
258 studies in other elasmobranchs (Merly et al., 2019). Intriguingly, arsenic levels were
259 substantially higher in the blood plasma than those found in muscle tissues of *S. lewini* (Bergés-
260 Tiznado et al., 2021; Boldrocchi et al., 2019), where arsenic would be expected to concentrate.
261 Hypothetically such a difference could result purely from differential local arsenic abundance
262 and consequently differential exposure, however similar relationships between blood and
263 muscle tissue concentrations of arsenic have also been reported in *Carcharodon carcharias*
264 (Merly et al., 2019). Exposure to arsenic in aquatic environments can have a range of damaging
265 physiological consequences even if only a relatively small proportion of arsenic is found in a
266 toxic form (Dringen et al., 2016; Greani et al., 2017), and for this reason we suggest that both

267 *S. lewini* and *C. carcharias* may possess arsenic detoxification mechanisms In the event that
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2 268 no such detoxification mechanism exists, the reported blood plasma arsenic (As)
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4 269 concentrations in both taxa would be extremely concerning and suggest that urgent
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7 270 modifications to local management plans are required. Selenium and zinc have both been
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10 271 shown to provide protection against arsenic toxicity in fishes (Roy and Battacharya, 2006; Zeng
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12 272 et al., 2005), however neither of these elements were present in particularly high concentrations
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14 273 (Table 1). Nickel (Ni) exhibited a similar trend, present at a concentration similar to that found
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17 274 in *C. carcharias* (Table 1; Merly et al., 2019) and significantly higher than that found in *S.*
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19 275 *lewini* muscle tissue (Boldrocchi et al., 2019). Whilst considered a potentially toxic element in
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22 276 many taxa (Blewett and Leonard, 2017), there is mounting evidence that Ni may be essential
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24 277 for fishes (Chowdhury et al., 2008; Pyle and Coutoure, 2011), and thus it remains to be seen
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27 278 whether arsenic and nickel blood plasma concentrations are driven by common underlying
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29 279 processes or not.

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34 281 Whilst some elements were found in far higher concentrations than expected based on studies
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36 282 utilizing muscle and liver tissues, others did not differ noticeably: copper and cadmium
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39 283 concentrations (Table 1) were found to be comparable to those reported by in *S. lewini* muscle
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41 284 tissue (Powell and Powell, 2001; Ruelas-Inzunza et al., 2020). Selenium (Se) is an intrinsic
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44 285 component of the central nervous system (Steinbrenner and Sies, 2013); however it can be
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46 286 toxic when present at high levels (Berges-Tiznado et al., 2015; Peterson et al., 2009; Ruelas-
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49 287 Inzunza et al., 2020). The Se levels in *S. lewini* blood plasma were found to be low (Table 1)
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51 288 in comparison with both other *S. lewini* tissues (Boldrocchi et al., 2019) and *C. carcharius*
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53 289 blood plasma (Merly et al., 2019). Mercury (Hg) concentration was also far lower than
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56 290 expected on the basis of existing studies (Berges-Tiznado et al., 2015; Merly et al., 2019).
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58 291 Given the toxic effects of both mercury and selenium at high concentrations (Escobar-Sánchez
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292 et al., 2010, 2011; Lemly, 2002) one could suggest that these results have positive implications
293 for *S. lewini* population-level health. However, this study utilized exclusively juvenile
294 individuals, and thus the trophic position of said individuals (and consequently the extent to
295 which biomagnification has occurred) is likely to be lower than in adult individuals utilized in
296 other studies (Cerutti-Pereyra et al., 2022; Estupiñán-Montaña et al., 2021).

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298 Whilst the absolute concentration of selenium was lower than expected in *S. lewini* blood
299 plasma, the Se:Hg ratio was found to be relatively high in most sampled individuals (Figure
300 4). This result is consistent with previous studies addressing mercury accumulation in *S. lewini*
301 (Berges-Tiznado et al., 2015; Pancaldi et al., 2019; Ruelas-Inzunza et al., 2020), and suggests
302 that Se is indeed performing a role in detoxification of Hg. This protective effect is thought to
303 be multifaceted, with competitive inhibition, formation of Hg-Se complexes and increase in
304 glutathione peroxidase activity (which inhibits oxidative damage by Hg) all thought to
305 contribute to the redistribution, detoxification, and excretion of mercury (Ralston et al., 2007;
306 Pancaldi et al., 2019; Raymond and Ralston, 2020).

307
308 Several correlations between the concentrations of different elements were detected: namely,
309 iron concentration was positively correlated to zinc and selenium concentrations, a trend which
310 has also been reported in *C. carcharias* (Merly et al., 2019) and other taxa (Schmitt et al.,
311 2007). Such correlations are thought to occur due to significant interplay between the
312 regulatory mechanisms governing the usage of these metals (Ehrensburger and Bird, 2011). It
313 has been reported that high levels of iron and zinc could potentially offset the toxic effects of
314 arsenic by replacement (Merly et al., 2019; Wang et al., 2020), however we are unable to verify
315 or refute such claims based on our results.

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317 Generally, our results demonstrate that trace element and heavy metal concentrations are higher
318 in the blood plasma of *S. lewini* individuals than from muscle and liver tissues of *S. lewini*
319 individuals reported in other studies (Table 1; Berges-Tiznado et al., 2015; Pancaldi et al.,
320 2019; Ruelas-Inzunza et al., 2020). We suggest two potential explanations for this trend.
321 Firstly, apparent high concentrations may result from the differences in biochemistry and
322 turnover rate between blood plasma and muscle/liver tissues in this species, as observed in
323 other taxa (Kim et al., 2012; Whitehead et al., 2020). Alternatively, the population sampled in
324 this study may be exposed to higher levels of trace elements and heavy metals through their
325 local environment. There is known to be significant variation in the chemical composition of
326 sediments and water column subdivisions in this region, due to wastewater discharge from the
327 city of La Paz, and nearby phosphate mining activity (González-Yajimovich et al., 2010; Páez-
328 Osuna et al., 2017). If the latter is responsible, this raises urgent questions regarding the health
329 and stability of *S. lewini* populations in this area. Whilst relatively high levels of heavy metals
330 have been found to be of minimal physiological influence in other taxa (Merly et al., 2019) and
331 we found evidence for detoxification mechanisms operating in *S. lewini*, no haematic
332 information regarding condition of health of sampled individuals was available. Thus, high
333 heavy metal concentrations could instead be symptomatic of population declines which have
334 been occurring in the region for years (Gallagher and Klimley, 2018; Pérez-Jiménez, 2014).
335 Further studies are urgently required to determine the extent to which *S. lewini* populations are
336 affected by the elemental concentrations reported in this study, and any indirect consequences
337 this might have for the wider ecosystem. Importantly, these hypotheses are not mutually
338 exclusive and both anatomical and local biogeochemical differences may contribute to
339 observed elemental concentrations.

1 341 A comparison of the concentration levels of element with maximum limits permitted for the
2 342 human consumption of a food products provided by the Food and Drug Administration and the
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4 343 World Health Organization (FAO Legal Notice No 66/2003) produced some concerning
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7 344 results. A total of 18 sampled individuals recorded Se concentration, higher than the limit
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9 345 established by the FAO at 1.0 µg/g, as well as one individual with mercury concentration (0.52
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11 346 µg/g) in excess of the limit established by the FAO for human consumption (0.50 µg/g).
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13 347 Additionally, Cd concentrations were also found to exceed the limit established in 2003 by the
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15 348 FAO and by the official Mexican regulation (NOM 242-SSA1, 2009) (0.05 ppm) in 40 out of
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17 349 55 sampled individuals. Whilst this would be concerning regardless of the population from
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19 350 which individuals were sampled, shark consumption in this region is high at 17.34 g per capita
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21 351 per day, which represents over 20 times the national average (Ruelas-Inzuna et al., 2020).
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23 352 Whilst many species of sharks are consumed in the region, *S. lewini* represents one of those
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25 353 most frequently available. In the absence of proportionate and effective conservation measures
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27 354 to protect *S. lewini* from overfishing in Mexican waters, measures limiting shark consumption
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29 355 or enforcing existing limits for heavy metal and trace element consumption are urgently
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31 356 required in the interests of public health.
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42 358 **Differences between sex and size:**

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46 360 We did not recover evidence of any statistically significant relationship between elemental
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48 361 concentrations and sex (Figure 3). This result is not unexpected and matches findings in other
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50 362 taxa of similar trophic position (Adel et al., 2017; Escobar-Sánchez et al., 2011; Merly et al.,
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52 363 2019). This suggests that elemental uptake and processing in *S. lewini* does not differ
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54 364 significantly between the sexes. Whilst sex-based and size-based differences in habitat usage
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56 365 and trophic ecology are known in this species (Estupiñán-Montaño et al., 2021; Hoyos-Padilla
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1 366 et al. 2014; Klimley, 1987;), the inclusion of only of juveniles in this study likely nullified
2 367 these potential confounding factors. On the basis of these results, we suggest that sex-based
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4 368 differences in elemental concentrations reported in other studies of *S. lewini* that include mature
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7 369 individuals result from differential exposure, rather than differential uptake and/or processing
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10 370 of said elements.

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14 372 Whilst no relationship between sex and elemental concentrations was recovered, we did find
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17 373 evidence of significant relationships between size and the concentration of several elements.
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19 374 Iron and chromium concentrations were both higher in subadult individuals than immature
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22 375 individuals (Table 2), and iron, chromium, copper, and zinc concentrations all showed
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24 376 significant positive relationships with total length (Table 3). These results differ somewhat
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27 377 from existing literature concerning ontogenetic trends in elemental accumulation in
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29 378 elasmobranchs, where Zn and Cr typically demonstrate negative relationships with body length
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32 379 (Boldrocchi et al., 2019; Endo et al., 2008). Comparatively low concentrations of zinc in
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34 380 smaller individuals are concerning given that physiological demand for zinc is thought to be
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37 381 greatest during early ontogeny (Endo et al., 2008), however it is important to recognize that the
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39 382 relatively high turnover rate of blood plasma compared to muscle and hepatic tissues (Kim et
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41 383 al., 2012; Whitehead et al., 2020) means that observed elemental concentrations are likely less
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44 384 related to growth and more transient in nature (Merly et al., 2019). Neither sex nor total length
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46 385 was found to correlate significantly with Se:Hg ratio, in agreement with studies concerning
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49 386 other tissues and organs of *S. lewini* (Berges-Tiznado et al., 2015). This suggests that the
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51 387 protective effects of selenium against mercury toxicity remain consistent between the sexes
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54 388 and throughout ontogeny.

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391 **Conclusion:**

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2 392 This study is the first to report the concentration of heavy metals and trace elements in the
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5 393 blood plasma of the scalloped hammerhead shark (*S. lewini*), and amongst the first to analyze
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7 394 elemental concentrations in the blood plasma of any elasmobranch. Individuals of *S. lewini*
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9 395 appear to possess higher concentrations of several trace elements and heavy metals than
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11 396 populations from other regions, with concentrations of some elements such as Cadmium
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13 397 exceeding the limit determined safe for human consumption by national and international food
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15 398 standards agencies. High elemental concentrations may reflect biochemical differences
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17 399 between blood plasma and other tissues; however, they may also be symptomatic of high levels
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19 400 of exposure triggered by anthropogenic activities such as mining. Whilst *S. lewini* appears to
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21 401 possess detoxification mechanisms against some pollutants such as mercury, it remains to be
22
23 402 seen whether marine pollution has significant physiological implications in this population. We
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25 403 also provide evidence of elemental accumulation through ontogeny, the nature of which differs
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27 404 from that reported in previous studies. There is much uncertainty underlying the relationship
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29 405 between different trace elements and heavy metals, and future studies are urgently required to
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31 406 assess the implications of these results not only for human health, but for the conservation
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33 407 prospective of *S. lewini* populations and the wider ecological community.
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417 **CRedit authorship contribution statement**

418
419 **Darren A. Whitehead:** Conceptualization, Methodology, Formal analysis, Investigation,
420 Resources, Writing – Original draft and editing, Visualization, **Joel H. Gayford:** Formal
421 analysis, Investigation, Resources, Writing – Original draft and editing, **Jacopo Gobbato:**
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423 Investigation, Writing – original draft, **Maria Tringali:** Formal analysis, Methodology,
424 Visualization, **James T. Ketchum:** Supervision, Project administration, Supervision, Funding
425 acquisition, **Felipe Galvan Magaña:** Writing – editing, supervision, **Davide Seveso:** Writing
426 – editing, supervision and **Simone Montano:** Conceptualization, Writing – editing,
427 supervision.

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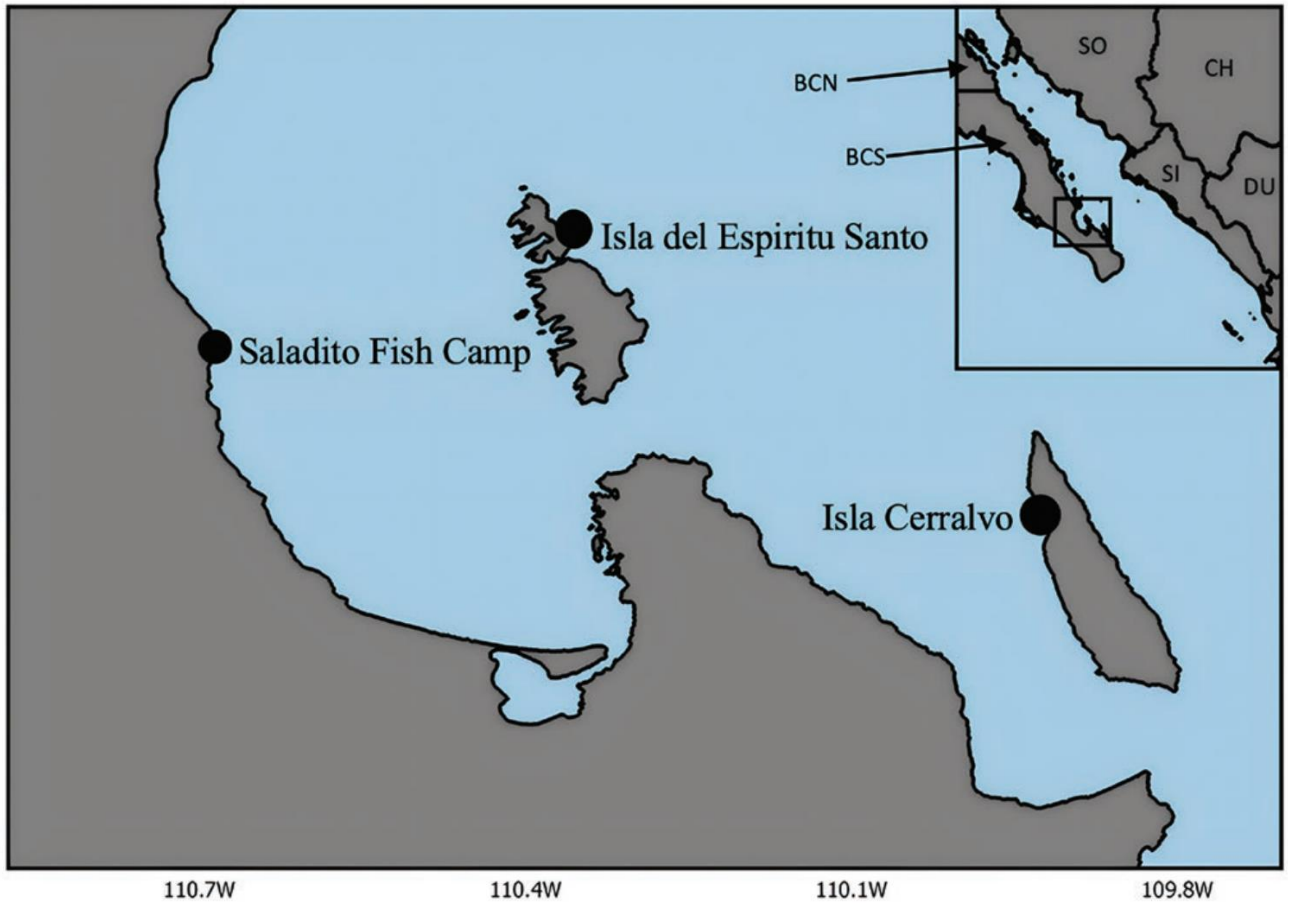
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759 **Figures & Tables**

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784 **Figure 1.** Map of the study site. Geographical labels on the inset refer to states of Mexico: Baja
785 California Norte (BCN), Baja California Sur (BCS), Sonora (SO), Sinaloa (SI), Chihuahua
786 (CH), Durango (D).

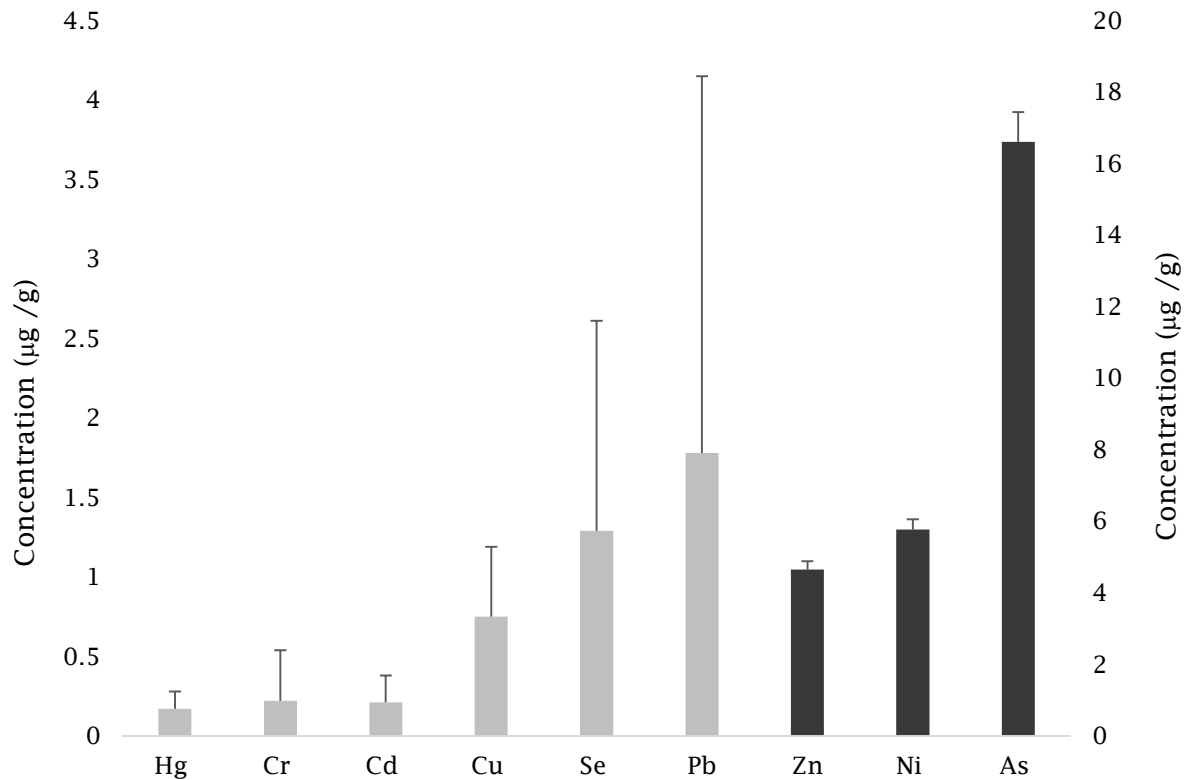


Figure 2. Concentration of all heavy metals and trace elements analyzed: mercury (Hg), lead (Pb), chromium (Cr), cadmium (Cd), copper (Cu), selenium (Se) zinc (Zn), nickel (Ni) and arsenic (As). Iron (Fe) is excluded from the figure as it has significantly higher concentration as key component of hemoglobin (Wells, 1999). Elements in grey corresponds to the left concentration scale, whereas the elements in black to the right concentration scale.

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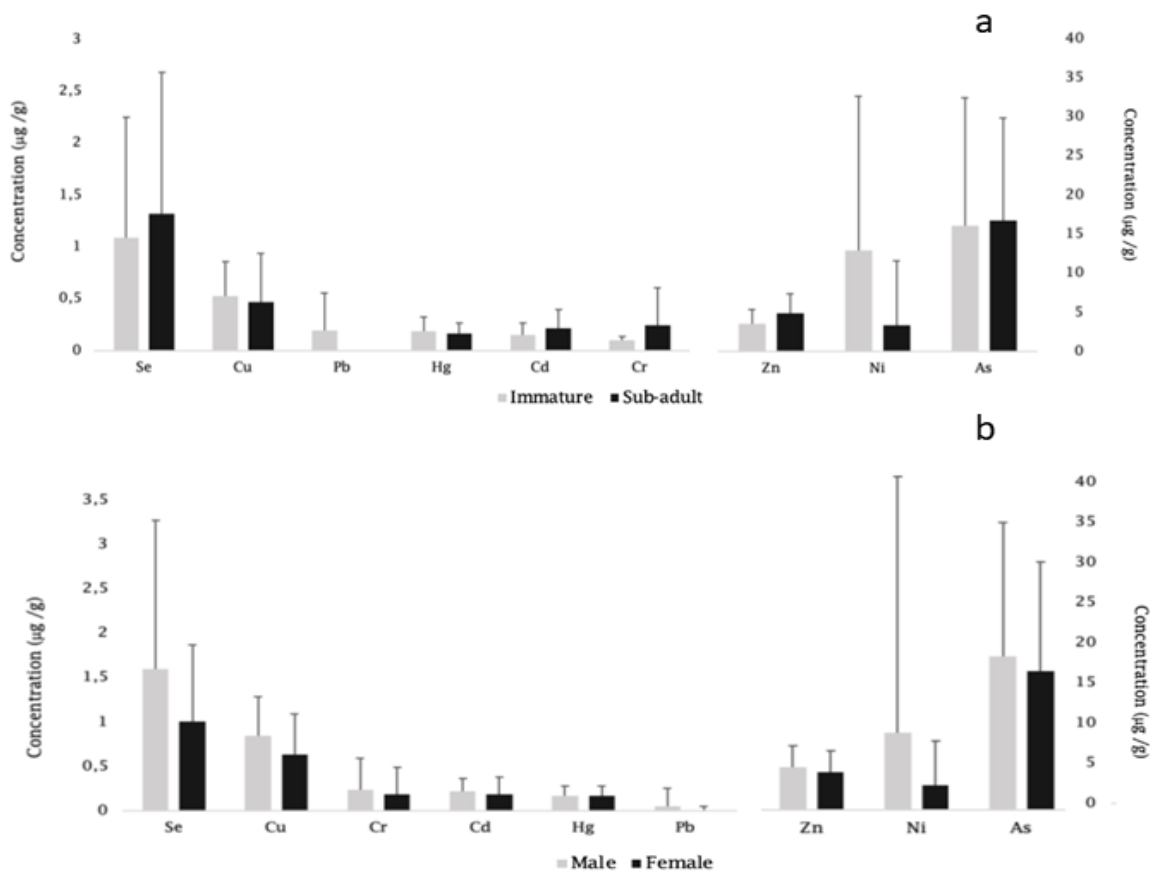


Figure 3. Mean elemental concentration: mercury (Hg), lead (Pb), chromium (Cr), cadmium (Cd), copper (Cu), selenium (Se) zinc (Zn), nickel (Ni) and arsenic (As) for (a) immature and subadult and (b) male and female.

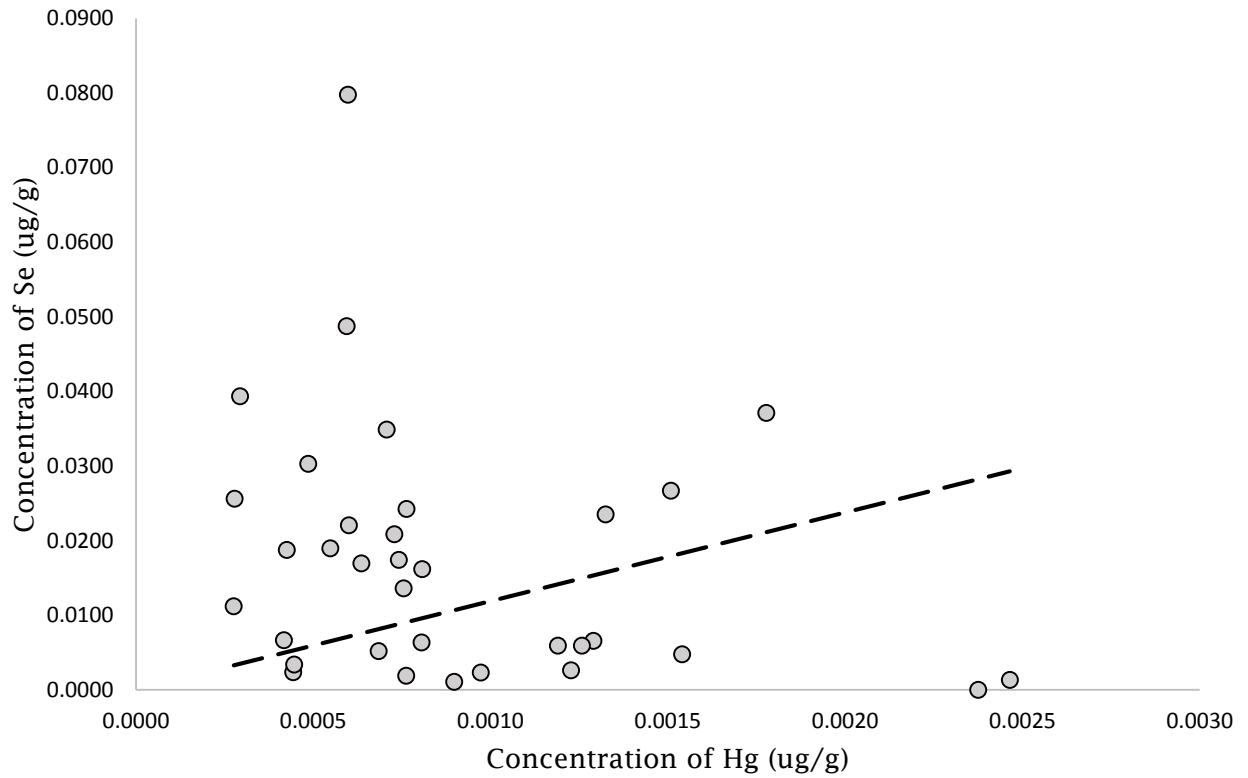


Figure 4. molar ratio between Se and Hg concentrations for sampled individuals.

Table 1. Overview of the concentration values of trace elements and heavy metals in the samples analysed. The limits for human consumption are displayed when assessed by FAO (Legal Notice No 66/2003).

Elements	Minimum (µg/g)	Maximum (µg/g)	Mean (µg/g ± SD)	Limit for human consumption (FAO) (µg/g)
Mercury (Hg)	0.06	0.52	0.17 ± 0.11	0.50
Selenium (Se)	0	6.30	1.29 ± 1.32	1.00
Zinc (Zn)	0	9.98	4.65 ± 2.38	30.00
Lead (Pb)	nd	5.26	1.78 ± 2.37	0.2
Arsenic (As)	0	16.61	16.61 ± 13.87	*
Iron (Fe)	0.81	918.60	459.56 ± 232.54	*
Chromium (Cr)	0	1.80	0.22 ± 0.32	*
Cadium (Cd)	0	0.85	0.21 ± 0.17	0.05
Nickel (Ni)	0	144.13	5.77 ± 20.60	*
Copper (Cu)	0	1.61	0.75 ± 0.44	*

* no limit established yet by FAO

869 **Table 2.** Statistically significance difference and correlation between concentration of Fe and
1 870 Cr in Immature and Sub-adult individuals.

Elements	Spearman's Rho Test	<i>P</i> value	Mann-Whitney Test	<i>P</i> value
Iron (Fe)	0.330	0.016	2.377	0.017
Chromium (Cr)	0.327	0.015	2.400	0.016

8 871
9 872
10 873 **Table 3.** Statistically significance correlation between elements concentrations and TL of the
11 analyzed individuals.
12 874

Elements	Spearman's Rho Test	<i>P</i> value
Iron (Fe)	0,339	0.013
Chromium (Cr)	0,564	<0.001
Zinc (Zn)	0,335	0.012
Copper (Cu)	0,322	0.020

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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: